Improving Water and Nutrient-Use Efficiency in Food Production Systems

.uc. Rectification

1 Current State and Future Potential of Global Food Production and Consumption

Christine Heumesser, Simon Thaler, Martin Schönhart, and Erwin Schmid

Introduction

The Food and Agriculture Organization (FAO) estimated the number of undernourished people in the world to be 925 million in 2010, which was 98 million below 2009 levels (FAO 2010c). Hence, more than 1 in 7 people live on a caloric intake below the minimum dietary energy requirement needed for light physical activity. However, the share of hungry people in the world has been declining since the mid-1990s and is at present below the 1970 level (FAO 2009c).

By 2050, global population is projected to reach 9 billion people (United Nations 2009). The continued population growth and the increasing per capita real income will further increase a total food demand for the next 40 years, with changing dietary patterns toward higher proportions of meat, dairy, and fish as well as processed food (Godfray et al. 2010). FAO (2009a) estimated that the current global food production needs to increase by 70% to meet the total food demand in 2050. On average, global agricultural production is projected to grow at 1.7% in the current decade, compared with 2.6% in the first decade of the 21st century (Organisation for Economic Co-operation and Development [OECD] and FAO 2011).

Meeting the increasing food demand is an unprecedented challenge. Even if attainable under the prospect of changing climate and decreasing growth rates of crop yields (Bruinsma 2003; Schmidhuber & Tubiello 2007), it will be difficult without severely exploiting and degrading natural resources, such as land, water, mineral nutrients, and fossil fuels. Additionally, the price hike of commodities and basic staples from 2006 onward and the subsequent financial and economic crisis from 2009 have drastically affected the number of people suffering from hunger and undernourishment (FAO 2009c). High commodity prices increased aggregated consumer price inflation, reduced purchasing power of poor populations, and negatively affected economic stability and food security (FAO-OECD 2011). For many developing countries, the global economic crisis led to a reduction in export earnings, remittances, foreign direct investment, and foreign aid, which led to employment and income losses (FAO 2009c). The price developments were driven by the connection between the agricultural and energy markets, increasing demands for cereals and oilseeds for biofuel production, weather-induced shortfalls of some food products, historically low grain stockpiles, a declining US dollar, increasing agricultural costs of production, and growing foreign exchange holdings by major food-importing countries (Trostle 2008).

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Food security does not only encompass food availability and supply, but it also includes food access (which is determined by political, social, and economic arrangements), food use, and food stability (FAO 2006). In this chapter, we focus on food availability and supply by investigating the current state of, and the future potential for, global, resource-efficient food production and consumption.

We first identify options and challenges in increasing global food production. This includes the expansion of agricultural land and competing usage paths (i.e., food, feed, biofuel, and nature conservation as well as increasing agricultural production by intensifying crop management). Furthermore, we discuss the impacts of changing climate and weather patterns on food production together with the options to decrease food demands by changes in human consumption behavior (i.e., less meat in the diet and reducing food waste). In addition, we provide an overview of the trends and challenges concerning the efficiency of water and nutrient use that will be a crucial factor in managing competing uses (i.e., food, feed, fiber, and biofuel) as well as negative environmental externalities.

Global Food Production

In this section we contrast frequently raised options and challenges to meet the increasing global food demand. We investigate the supply side of the global food production, focusing on the expansion of agricultural land and the productivity growth, in particular through use of fertilizers, irrigation, and biotechnology. We also account for climate change as an overarching challenge, affecting the future production strategies.

Agricultural Land Expansion

The world's total land area amounts to approximately 13 billion ha, of which approximately 5 billion ha (38.5%) are agricultural land. Of that land only 1.4 billion ha (28.6%) are arable land (FAO 2010a). Historically, the expansion of agricultural land has been a way to meet the rising food demand. From the 1960s onward, however, food production has been decoupled from cropland expansion as a result of considerable productivity increases (Lambin et al. 2003). Between the early 1960s and the late 1990s, arable land and land under permanent crops expanded by 155 million ha, or 11%, while world population almost doubled. Arable land per person fell by 40% from 0.43 ha to 0.26 ha on average, but land productivity growth through intensification compensated for this reduction in area per person (Bruinsma 2003).

To meet the increasing food demand, a remaining question is whether further expansions in agricultural land are necessary as well ecologically and socioeconomically feasible.

Drivers of Land Use Change

The causes of land use change and agricultural land expansion are manifold and complex, involving situation-specific interactions among a large number of factors at different spatial and temporal scales (Geist & Lambin 2002; Lambin et al. 2003; Smith et al. 2010). Lambin et al. (2003) identified five high-level causes of land use change: (1) resource scarcity and related pressures on natural resources, (2) changing market opportunities, (3) outside policy interventions, (4) loss of adaptive capacity and increased vulnerability of local land users, and (5) changes in social organizations, institutions, and human attitudes. Also, Smith et al. (2010) identified socioeconomic, technological and institutional factors, and social trends, such as population growth and urbanization, as the

underlying causes for competition for land. These factors determine the extent of direct pressures on land, which include land transition (e.g., forest clearing to grow crops and pastures), land degradation (e.g., logging, induced fires or overgrazing), and natural causes (e.g., climate change and water availability).

Growing crops for bioenergy has been identified as potential competitor to food production, potentially risking the displacement of forests and grasslands through direct and indirect land use change effects. The International Energy Agency (IEA) estimated that energy crop production took place on 1% of global arable land in 2004 and may increase to between 2% and 3.8% by 2030 because biofuel legislation in several countries supports its expansion. The European Union and North America are predicted to experience the largest growth in the area under biofuel crops, from 1.2% of arable land in 2004 to 11.6% in 2030 (EU), with corresponding numbers for the United States of 1.9% to 5.4%. Comparatively, the land requirements for biofuel production in other parts of the world might increase from 0.1% to 2.7% of arable land in 2030 (IEA 2006). The degree of competition can be reduced by technological progress in biofuel conversion technologies and a switch to second-generation technologies using agricultural and forestry by-products (Fischer et al. 2009).

Land degradation and the subsequent loss of productive capacity could potentially lead to an expansion of agricultural land into remaining natural habitats. Land degradation is increasingly driven by improper agricultural land use, poor soil and water management practices, deforestation, loss of natural vegetation, or excessive use of agro-chemicals, as well as, natural disasters including droughts, floods, and landslides (United Nations Environment Program [UNEP] 2002; Bruinsma 2003). According to various global land degradation assessments (Oldeman et al. 1990; UNEP 1992; Bridges & Oldeman 2010; FAO 2012), approximately 23% of all usable land (excluding mountains and deserts) has been affected by degradation to a degree sufficient to reduce its productivity. In the early 1990s, about 910 million ha of land were classified as "moderately degraded" with greatly reduced agricultural productivity and 305 million ha were classified as "strongly to extremely degraded" (UNEP 2002).

The expansion of agricultural land contributes to the loss of natural ecosystems and corresponding biodiversity losses (Koh & Ghazoul 2008). Nellemann et al. (2009) estimated that 80% of all endangered birds and mammals are threatened by agricultural expansion and unsustainable land use. In a majority of developing economies, the decline in forest and woodland area is mainly the result of land conversion to crop production (FAO 2007).

The Global Forest Resources Assessment 2000 estimated deforestation during the 1990s at 16.1 million ha per year, resulting in a loss of 4.2% of the natural forest that existed in 1990 (FAO 2001). In the period from 1981 to 1990, the area of tropical forests cleared each year in Latin America was 7.4 million ha on average. This is almost as much as the sum of deforested areas in Asia and Africa combined. During 1991 to 2000, deforestation in Latin America declined to 4.3 million ha annually (Barbier 2004). At the same time, there was an increase in the forest area as a result of aforestation, such that the net global decrease in forest area was about 9.4 million ha per year from 1990 to 2000. Overall, the total net forest change was positive for the temperate regions but negative for the tropical ones (FAO 2001).

Deforestation has various adverse effects. In 2004, carbon dioxide-equivalent emissions from deforestation, decay of biomass, and burning of peat land were estimated to be 17.3% of total emissions (International Panel for Climate Change [IPCC] 2007). In addition, tropical forests are rich in floral and faunal diversity, which is threatened by deforestation. Even though a slowdown in deforestation and rangeland clearance for crop production has been observed on a global scale, pressures on forests are likely to continue in some developing countries. Deforestation is driven by a number of site-specific causes (Geist & Lambin 2002), such as a lack of nonagricultural employment

opportunities in a large proportion of rural communities (Bruinsma 2003) or unfavorable management options (e.g., low irrigation and fertilizer rates leading to soil degradation and, consequently, expansion of agricultural lands) (Barbier 2004).

A major driving factor for deforestation is the expansion of grazing land for livestock, particularly in Latin America. About 70% of deforested land in the Amazon is now managed as pastures. On a global scale, the livestock sector is estimated to account for 78% of agricultural land and as much as 33% of the cropland. Dietary shifts toward more meat will require a much larger share of crop and grazing land for feed production, which will exert pressures on crop production for human uses (Steinfeld et al. 2006).

The literature is ambiguous on whether further agricultural land expansion is feasible or not. For instance, researchers from FAO and the International Institute for Applied Systems Analysis (IIASA) developed an agro-ecological zones model that computes amounts of nonarable and arable land as a function of environmental constraints (Fischer et al. 2002a, b; Fischer et al. 2005). Fischer et al. (2002b) estimated that approximately 2.5 billion ha or 18.6% of land with a potential for rain-fed crop cultivation exist, of which 1.6 billion ha are located in developing countries. This estimate does not account for nonagricultural land uses such as infrastructure, settlements, or legally protected areas. However, even when excluding areas that are forests or legally protected areas, 17.6% of total terrestrial surface has a potential for arable uses (Fischer et al. 2002b). On a regional scale, it is suggested that only 22% of potentially suitable arable land in sub-Saharan Africa, 19% in Latin America, and 52% in East Asia (excluding China) was farmed from 1997 to 1999. In these regions, expansion of arable land continues to contribute to agricultural growth. In contrast, about 87% of suitable area has already been cultivated in the near East and North Africa and about 94% in South Asia (except India) (Bruinsma 2003).

Other studies conclude that much of the land suitable for agricultural production has already been developed (Khan & Hanjira 2008). For instance, Alexandratos (1995) estimated that more than 70% of the potentially available rain-fed cropland in sub-Saharan Africa and Latin America suffers from topographical, soil, and terrain constraints and therefore is not available for agricultural production. Other models predict that under severe climate change, the global amount of land suitable for agriculture will remain the same in 2080 as it was in the early 2000s (Fischer et al. 2002a, b; Parry et al. 2004). Also, Fischer et al. (2002b) concluded that there were severe limitations to their estimates of land with potential for arable uses. An increased use of cultivated land might not be feasible because of competition for land with alternative agricultural uses or severe impacts on biodiversity and the global carbon cycle. Additionally, there might be ecological constraints, low soil fertility, high incidence of crop diseases, or a lack of infrastructure and access to appropriate technologies (including economic incentives to adopt them). Socioeconomic restrictions (e.g., suitability for a particular crop that is not demanded on the domestic or foreign markets) are further limitations to the estimates of potentially available arable lands (Fischer et al. 2002b; Bruinsma 2003).

Overall, some additional land could be used for crop production, but the competition with other land uses, the desire to protect natural habitats, and the required services provided by natural ecosystems (e.g., carbon storage in rainforests and flood control) can make this an unwanted or inefficient solution (Balmford et al. 2005).

Productivity Growth

Intensification can be defined as an increase in production per unit of inputs (e.g., labor, land, time, fertilizer, seed, feed, or cash) (FAO 2004). It has permitted the doubling of the world's food

production from 1961 to 1996, with only a 10% increase in the global amount of land under cultivation (Tilman 1999). Driving forces for intensification are releasing capital and knowledge constraints, changes in the price ratios of inputs and outputs, as well as farm technologies as a function of land and water scarcity, growth in population, and investments in crop and livestock breeding that can change the quantity and value of production per ha (Lambin et al. 2001).

Even though yield growth is expected to remain the driving force of crop production, annual yield growth rates for many crops are projected to decline until 2030. The average increase in cereal yield in developing countries has declined from 2.5% per acre in the period from1961 to 1999 to 1.4% between 1991 and 2001 and is projected to decline to 1% in 2030 (Bruinsma 2003). However, there is a wide geographic variation in crop productivity, even across regions with similar natural conditions, for example because of inadequate nutrient and water management (World Bank 2008; Vitousek et al. 2009). There are institutional constraints such as limited access to knowledge and technologies to increase production, lacking finances to undertake investments (e.g., irrigation, fertilizer, soil-conservation measures), and unfavorable prospects for returns on agricultural investment. Closing the yield gap (i.e. the difference between realized productivity and the maximum attainable yield at a site given current genetic material, available technologies, and management) can substantially increase food production levels (Godfray et al. 2010; Foley et al. 2011).

The gains in agricultural productivity are often accompanied by adverse effects on natural resources and the environment, which may risk the future productive potential. Examples are land degradation through soil erosion and salinization; susceptibility to diseases; loss of genetic resources; emissions of greenhouse gases; nitrogen and phosphorus losses causing eutrophication in water aquifers; or losses of habitat and species diversity (Bruinsma 2003).

Increasing land-use intensity such as through higher fertilizer inputs and mowing frequencies as well as homogenization of landscapes reduces biodiversity (Benton et al. 2003). Globally, more than 4,000 plant and animal species are threatened by agricultural intensification (Nellemann et al. 2009). There is an ongoing debate whether the land for nature protection should be separated from the agricultural land use. It may be beneficial under certain circumstances to intensify production in some areas to reduce pressure on the native lands for nature conservation (Balmford et al. 2005; Green et al. 2005). However, one has to acknowledge that leakage and rebound effects may undermine expected land-use effects (Lambin & Meyfroidt 2011). On the other hand, for some parts of the world, such as the European Alps, extensively managed agricultural land-use systems have created semi-natural habitats of high ecological value, within which both intensification as well as land abandonment may lead to biodiversity losses (Tasser & Tappeiner 2002).

Fertilizer Use

One-third of the increase in world cereal production in the 1970s and 1980s has been attributed to increased use of fertilizers (Bruinsma 2003). The other estimates based on the FAO database (FAO 2010a) suggest that between 1961 and 2007, the use of nitrogen fertilizer on a global scale increased 7.5-fold and that of phosphorus 3.3-fold (see Table 1.1).

About 40% of the global human population is dependent on synthetic nitrogen fertilizer (Smil 2002a; Stewart et al. 2005). However, its use varies among regions (Vitousek et al. 2009). In 1997 to 1999, the highest rates of fertilizer use were in East Asia (194 kg/ha of arable land), followed by the industrial countries with 117 kg/ha, whereas farmers in sub-Saharan Africa applied only 5 kg/ha on average (Bruinsma 2003). This resulted in average cereal yields in sub-Saharan Africa of 1.1 t/ha in 2000, whereas average yields in Asia, Latin America, and the Middle East/North Africa amounted to 3.7, 2.8, and 2.7 t/ha, respectively (Kelly 2006).

	Nitrogen fertilizers (total N) in kg/ha					Phosphate fertilizers (total P ₂ O ₅) in kg/ha				
	1961	1981	1997	2002	2007	1961	1981	1997	2002	2007
Sub-Saharan Africa	0.5	3.3	3.7	2.7	2.6	0.3	2.3	2.3	1.5	1.5
Near East/North Africa	3.9	27.9	45.5	50.7	62.8	1.3	14.0	17.7	19.5	20.4
South Asia including India	1.8	26.0	68.8	69.2	89.7	0.4	8.5	22.7	24.7	32.5
South Asia excluding India	3.4	35.0	91.7	105.1	112.0	0.8	11.2	21.1	29.1	32.4
East Asia including China	5.8	77.5	121.8	143.0	150.7	2.5	20.9	47.9	48.6	57.0
East Asia excluding China	6.6	33.1	53.8	54.9	60.9	4.0	11.0	18.5	17.0	21.3
Latin America and the	4.2	20.1	29.5	29.8	43.3	3.5	15.3	20.5	24.6	37.1
Caribbean										
Industrial countries	16.3	50.8	57.1	61.2	65.2	18.0	28.1	24.3	24.8	24.9
Transition countries	18.7	85.8	18.3	16.0	20.5	10.9	56.4	4.9	5.2	7.5
World	8.5	41.5	53.5	56.4	64.4	8.0	21.3	21.9	22.2	26.3

Table 1.1 Nitrogen and phosphate fertilizer in kg/ha arable land and permanent crops for the years 1961, 1981, 1997, 2002, and 2007.

Note: The country groups correspond to the classification proposed in Bruinsma (2003).

Source: FAO (2010a).

In the past decades, an increase in the consumption of nitrogen and phosphorus fertilizers has been observed globally (see Table 1.1). By 2050, nitrogen fertilization is expected to increase by 2.7 times and phosphorus by 2.4 times on a global scale (Tilman et al. 2001). However, increased fertilizer application rates exhibit diminishing marginal returns such that further increases in fertilizer are unlikely to be as effective in increasing cereal yield as in the past. A declining trend in global nitrogen efficiency of crop production (annual global cereal production divided by annual global nitrogen application) is shown in Figure 1.1 (Tilman et al. 2002; FAO 2010a).

It is estimated that today only 30% to 50% of applied nitrogen fertilizers (Smil 2002a; Ladha et al. 2005) and 45% of phosphorus fertilizers (Smil 2000) are used for crops. For example, only 20% to 60% of nitrogen fertilizers applied in intensive wheat production is taken up by the crop, 20% to 60% remains in the soil, and approximately 20% is lost to the environment (Pilbeam 1996). The phosphorus-use efficiency can be as high as 90% for well managed agroecosystems (Syers et al. 2008) or as low as 10% to 20% in highly phosphorus-fixing soils (Bolland & Gilkes 1998).

Energy is the most important resource for the production of synthetic nitrogen fertilizers. This is not the case for phosphorus fertilizers, which are produced from nonrenewable phosphate minerals. According to Cordell et al. (2009), peak phosphorus could occur by 2033. Because 90% of the rock phosphate is used for food production (Smil 2002a), a responsible handling is crucial to meet the future demand.

Even if the resource scarcity can be managed properly, there are potential adverse impacts on the environment. In regions with excessive use of fertilizers, the main environmental concerns are the off-site effects of losses to the atmosphere, as well as, surface and ground waters. The major environmental impact of phosphorus fertilization is eutrophication of surface waters. For nitrogen fertilization, these environmental impacts include eutrophication of coastal seas and lakes, loss of biodiversity and changes in species composition in terrestrial and aquatic ecosystems, groundwater pollution with nitrate and nitrite, and increasing greenhouse gas emissions. Furthermore, direct toxicity of inorganic nitrogenous compounds can severely threaten aquatic ecosystems (Smith et al. 1999; Bennet et al. 2001; Tilman et al. 2002; Rabalais 2002; Camargo & Alonso 2006).



Figure 1.1 Diminishing returns of nitrogen fertilizer application. *A*, Global cereal production (tonnes/ha). *B*, Global nitrogen fertilizer efficiency of cereal production (annual global cereal production in tonnes divided by annual global nitrogen fertilizer production in tonnes for domestic use in agriculture.) *Sources*: FAO 2012a and Tilman et al. 2002.

Irrigation

Almost 97% of the global water resource is saltwater, 2% is snow and ice, and only 1% is available as liquid freshwater. Freshwater is mainly available as groundwater aquifers, whereas less than 2% is in rivers and lakes. Only a fraction of that is accessible for human use: an estimated 9,000 to

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14,000 km³ of water is economically available each year. This represents at most 0.001% of the estimated global water resources. Agriculture is the principal user of all water resources, accounting for 70% of all withdrawals (e.g., rainfall, water from rivers, lakes, and aquifers). In comparison, 10% is assigned to domestic uses and 20% to industrial uses (FAO 2003).

On a global scale, irrigation water represents only a fifth of the total water use in agriculture (Siebert & Döll 2009), although irrigation represents up to 95% of all water withdrawals in several developing countries, which plays a major role in food production and security (Siebert et al. 2006). In sub-Saharan Africa more than 95% of the farmed land is rain-fed, in Latin America 90%, in South Asia 60%, and in the near East and North Africa 75% (FAOSTAT 2005 cited in Wani et al. 2009). However, rain-fed agriculture remains risky because of spatial and temporal variability in rainfall, water scarcity, droughts, soil erosion by wind and water, low investment, or high population pressure (Wani et al. 2009). There is a significant difference in crop yields between irrigated and rain-fed agriculture. In developing countries, grain yields from rain-fed agriculture are 1.5 t/ha compared to 3.1 t/ha from irrigated agriculture on average (Rosegrant et al. 2002).

Although there is interest in increasing the productivity of rain-fed agriculture (Turral et al. 2010), a rise in irrigated agriculture is expected in the future. In developing countries, intensively managed irrigated agriculture already accounts for 40% of total crop production from just 20% of the arable land. The share of total crop production from irrigated agriculture is expected to increase to 47% by 2030 (Bruinsma 2003).

On the downside, intensive irrigation is considered a threat to the environment, causing salinization of soils that reduces soil fertility and, hence, food production. Salinization currently affects about 10% of the world's irrigated land (Schoups et al. 2002).

Finally, the question remains whether there will be sufficient freshwater resources to satisfy the agricultural and nonagricultural water demands (including protection of natural ecosystems) in the light of water pollution and largely unknown impacts of future climate change (Bruinsma 2003; FAO 2003). About 25% of the world's irrigated agricultural systems have been withdrawing water above the regeneration rate, thus creating a potential for serious shortages on a local to regional scale. This trend is aggravated by poor property rights specifications on water resources and inefficient irrigation practices resulting in land degradation and reduced productivity (FAO 2008). According to the World Water Assessment Programme (2009), a few countries have the data on the purposes, quantities, and qualities of water extraction as well as the remaining resources.

Even though water scarcity is difficult to measure, analyses indicate that more than half of the world's population lives in countries with varying degrees of water scarcity (FAO 2003). Water may be physically limited in the sense that there is no more water available, or economically, if a country cannot afford to develop additional water resources and infrastructures. It may also be caused by a lack of social adaptive capacity that prevents increasing productivity per unit water consumed (Allan 1995).

The water shortage on a global scale may be arguable. Some sources consider there is no serious threat of water shortages in the future (Bruinsma 2003), whereas others estimate that, taking environmental water requirements into account, serious water shortages are likely to occur (Smakhtin et al. 2004a, b).

Biotechnology

Plant breeding has created modern crop and animal varieties and contributed to productivity gains in agriculture. In the future, it may contribute to producing crops that are better adapted to pests, droughts, and other environmental stresses and have a higher nutritional value. In particular, genetically modified (GM) crop species such as maize, soybean, and cotton have been commercialized in the last decade, albeit in selected countries only. In 2007, GM crops were grown on estimated 114 million ha, with more than 90% in only four countries (Argentina, Brazil, Canada, and the United States) (World Bank 2010). The GM crop area expanded to 125 million ha in 25 countries, including 15 developing countries in 2008 (James 2008) and to 160 million ha in 29 countries in 2011 (James 2011).

The GM crops are expected to lessen environmental pollution, increase crop productivity, decrease production costs, and reduce nitrous oxide emissions (World Bank 2010). However, the cultivation of GM crops remains controversial. There are ethical concerns about increasing dependency of farmers on international biotechnology companies. Furthermore, concerns about food safety and environmental risks (including possible cross-pollination with wild relatives, creation of herbicide-tolerant weeds, and evolution of new pest types that are adapted to GM plants) are among the concerns. There are also fears that diversity of agricultural crop species may further decrease as traditional cultivars are displaced by a narrow range of new GM ones (World Bank 2010).

Climate Change

Climate change poses serious challenges to agricultural production and calls for alternative strategies to meet increasing food demands. At the same time, the agricultural sector contributes considerably to climate change. In 2004, the agricultural sector caused about 13.5% of global greenhouse gas emissions in terms of carbon dioxide-equivalent emission, including (1) nitrous oxide (N₂O) from fertilizers, (2) methane (CH₄) from livestock, rice production, and manure storage as well as (3) carbon dioxide (CO₂) from burning biomass. Emissions associated with the forestry sector, including land-use change, deforestation, and burning account for about 17.4% of total CO₂-equivalent greenhouse gas emissions (IPCC 2007; World Bank 2010).

Climate change is expected to put conflicting pressures on agricultural production (World Bank 2010). Although it is difficult to quantify the effect of climate change on food production compared to other drivers, the occurrence of following impacts is considered to be likely.

- (1) Increasing temperatures and changes in timing and volume of precipitation are likely to increase water and heat stresses on crops and livestock, particularly in subtropical regions. Even without reduced average precipitation, increased water stress can occur as high temperatures increase evapotranspiration.
- (2) An increase in the occurrence and intensity of extreme weather events (such as droughts and floods) can lead to reduced crop yields on average.
- (3) These factors can increase the risk of soil erosion, runoff, landslides, and pest invasions, all of which can threaten agricultural production (Padgham 2009).

Climate change impacts are predicted to vary across regions. For countries in the higher latitudes (such as Kazakhstan, the Russian Federation, or Ukraine), the food production potential could increase (Godfray et al. 2010). For instance, an increase in local annual temperatures of 1° to 3 °C along with enhanced CO_2 fertilization (World Bank 2010) and the rainfall changes may have beneficial effects on crop yields (Fischer et al. 2002a). To take advantage of these developments, the respective countries will have to invest in institutions and infrastructure (Fay et al. 2010). However, extreme weather events can offset, or even reverse, such beneficial impacts. For instance, when the increased likelihood of extreme weather events is taken into consideration for Russia, the years with

food production shortfalls are projected to triple by the 2070s (Alcamo et al. 2007). In contrast, in the low-latitude regions, even moderate temperature increases of 1° to 2°C can reduce yields of major cereals such as rice, wheat, maize, and sorghum (Easterling et al. 2007). These shortfalls can be fully or partially compensated for by implementing adaptation measures. However, if temperatures continue to increase, productivity could decrease significantly and the effectiveness of adaptation measures could even be eliminated (Padgham 2009).

Several adaptation strategies are suggested to buffer adverse climate change effects. They include rural livelihood diversification leading to increased economic security and less reliance on climate-sensitive agricultural activities (Padgham 2009) (e.g., income diversification, or diversification toward livestock, horticulture, and specialized agricultural production). It also involves increasing the genetic variability among individual crop varieties (World Bank 2010) and improving access to new varieties and other production technologies, which can help farmers to increase productivity and better manage risks from droughts and floods (Padgham 2009).

In addition to adaptation measures, mitigation strategies are required as well, such as improved crop, grazing, and livestock management. This can include agronomic practices such as efficient nutrient use, reduced tillage, and recycling management (IPCC 2007). Agricultural practices relying on the traditional knowledge of farmers might embody a wealth of location-specific adaptation and risk management options and aim to protect soils, biodiversity, and water aquifers (International Assessment of Agricultural Knowledge, Science and Technology for Development [IAASTD] 2009; World Bank 2010). Finally, policies and institutions need to be strengthened to encourage timely implementations of adaptation and mitigation practices (Padgham 2009).

Global Food Consumption

In this section, we present demand-side options to meet the increasing food demand, such as shifts in the dietary composition or reducing food wastage.

The world population is projected to rise to 9 billion people by 2050. The population in developing countries is projected to rise from 5.6 billion in 2009 to 7.9 billion in 2050, whereas population in developed countries is expected to increase from 1.23 billion to 1.28 billion (UN 2009). The increase in incomes of a large fraction of the world's population may be accompanied by substantial increases in consumption of food as well as quantities of waste or discarded food (Henningsson 2004).

Meeting the world's food requirements in the future may require changes in dietary patterns and composition as well as a reduction in food waste. Food intake in terms of calories, proteins, and fat differs considerably between developed and developing countries (Table 1.2),

	Energy consumption kcal/capita per day	Protein consumption g/capita per day	Fat consumption g/capita per day
World	2,770	76	78
Developed countries	3,380	102	129
Developing countries	2,620	70	66

Table 1.2 Dietary patterns per capita in developed and developing countries as well asworldwide (2003–2005).

Note: The values include household food wastes. Categorization based on SOFI Country Group Composition.

Source: Food and Agriculture Organization (2010b).

reflecting the mismatch between overconsumption of food in some regions of the world and undernourishment in others.

The physiological energy demand of humans depends on a number of factors such as age and sex, as well as, physical constitution and activity. However, assuming frequently stated daily average energy consumption of 2,500 kcal/capita, the average energy consumption in developed countries appears rather high in contrast to the group of developing countries. Such excess or "luxury" demand may consist of products with little nutritional value (such as soft drinks) but may have large impacts on resource demand and environmental quality. For example, the growth in calorie availability of about 600 kcal/capita per year from 1983 to 2000 in the United States has been estimated to result in an additional land demand of 0.46 ha/capita (Blair & Sobal 2006). Obesity and its corresponding health problems may be a consequence of excessive calorie intake; there are more than 1 billion people in developed and developing countries classified as obese (World Health Organization [WHO] 2003), leading to a double burden of both undernourishment and overconsumption in many countries (Schmidhuber & Shetty 2005).

Besides reductions in total food consumption in some parts of the world, shifts in the composition of diets from animal- to plant-based components are frequently seen as a further option to increase total food availability. Consumption of livestock products increased considerably during the previous decades as a consequence of declining relative prices as well as increasing incomes in developing and developed countries (Steinfeld et al. 2006; FAO 2009b). Today, the share of animal-based proteins in average diets is more than 60% in the United States, Germany, or Australia, but below 10% in countries such as Malawi, Burundi, or Mozambique (FAO 2010b).

As a result of conversion losses of feed in the animal metabolism, livestock products require a higher amount of land, water, energy, and nutrients than plant-based food to result in the same amount of calories and proteins. Only 3% to 20% of the plant-based energy and protein is converted to meat according to Smil (2002b). A complete shift from livestock toward plant-based diets would save about one-third of the world's cereal production used for animal feeding today (Steinfeld et al. 2006). However, any such estimates have to acknowledge the different nutritional values of livestock- and plant-based food products and the additional area necessary for plant-based food production. Furthermore, not all livestock is fed edible plant products but instead may use industrial by-products or forage from grasslands. Increases in the conversion efficiency from animal feed to human food (e.g. through better breeding or shifts toward more efficient alternative animal species, such as insects) can increase conversion efficiency (Godfray et al. 2010; Vogel 2010). Finally, a reduction in consumption of livestock products in developed countries may not necessarily lead to a decreasing livestock demand on a global scale because of compensation effects in developing countries (Rosegrant et al. 1999).

The reduction in food waste is a third demand-side option for increasing food availability. Godfray et al. (2010) summarized literature on food waste, showing global average values of 30% to 40% of food lost in both developed and developing countries. In developing countries, food losses mainly occur in the first steps of the food supply chain. This includes postharvest losses as a result of inappropriate storage, transport, and processing facilities as a consequence of inadequate technology, infrastructure, and management (Lundqvist et al. 2008; Godfray et al. 2010). In developed countries, losses occur mainly in the second step of the food chain (roughly 50% within households) (Godfray et al. 2010), where food (even eatable one) is discarded (Lundqvist et al. 2008). Therefore, without additional resource use, the food demands can be met by reducing the losses to a minimum.

Water- and Nutrient-Use Efficiency in Agriculatural Production

Efficient agricultural production and management practices are required to meet the growing demand for food without compromising the environment and agricultural resources. An important aspect is efficient water and nutrient use (i.e., increasing production per unit of added nitrogen, phosphorus, and water). However, there is diminishing marginal productivity associated with nutrient additions such that marginal crop yields are declining at higher fertilization levels (Tilman et al. 2002). Considering the nitrogen transfer from mineral fertilizer into the food eaten by humans, only 4% of the fertilized nitrogen can be found in animal-based and 14% in plant-based foods (Galloway & Cowling 2002).

A variety of practices and improvements are suggested in the scientific literature to increase nutrient-use efficiency in agriculture, such as the adoption of multiple cropping systems, improved crop rotations, or intercropping. Agroforestry, which includes trees in a cropping system, may improve pest control and increase nutrient- and water-use efficiency. Also, cover crops or reduced tillage can reduce nutrient leaching. Nutrient-use efficiency is increased by appropriately applying fertilizers and by better matching temporal and spatial nutrient supply with plant uptake (Matson et al. 1997; Tilman et al. 2002). Applying fertilizers during periods of highest crop uptake, at or near the point of uptake (roots and leaves), as well as in smaller and more frequent applications have the potential to reduce losses while maintaining or improving crop yield quantity and quality (Matson et al. 1996, 1997; Cassman et al. 2002). However, controlled release of nitrogen (e.g., via using nitrogen inhibitors) or technologically advanced systems such as precision farming appear to be too expensive for many farmers in developing countries (Singh 2005).

Many of the aforementioned management practices can be supported by targeted research (e.g., on improving efficiency and minimizing losses from both inorganic and organic nutrient sources; on improvements in timing, placing, and splitting of fertilizer applications, as well as by judicious investments, for example, in soil testing). Extension services (including programs that work with local and international seed companies to promote appropriate agronomic practices) are important as well (Cassman et al. 2002). Sound policies also include measures that make fertilizers and irrigation water available and are appropriately priced (Dorward et al. 2004). With proper incentives for sustainable fertilizer and water management, many countries could maintain agricultural growth and even reduce emissions and water pollution (World Bank 2010). In addition to improvements in crop production, the livestock production systems offer a large opportunity for enhancing nutrient efficiencies. Changing feed composition, optimizing manure management, and recycling animal waste have the potential to reduce nutrient losses substantially and therefore increase nutrient-use efficiency in agricultural production systems (Cowling & Galloway 2002).

Sustainable agricultural production also depends on making better use of existing water resources, for instance by increasing the efficiency of irrigation systems. Water efficiency usually refers to the percentage of available water (rainfall, surface water, or groundwater) that is used by plants (Wallace & Batchelore 1997). In contrast, irrigation efficiency describes the ratio between the estimated consumptive water use in irrigation and total irrigation water (including withdrawals). Irrigation water withdrawal usually exceeds consumptive water use because water is lost during storage, transport, and distribution from the source to the crops. In 1997 to 1999, irrigation efficiency was estimated at 38% in developing countries. In water-abundant areas (e.g., Latin America), it only reached 25%, whereas in water-scarce regions (such as the near East/North Africa or South Asia), it reached 40% to 44% (Bruinsma 2003). Some of the water lost may return to

aquifers from which it can be used again as long as infrastructural requirements are fulfilled and the water is of specific quality.

In rain-fed agriculture there are no losses as a result of storage or transport, but water is lost by runoff, soil evaporation, and percolation. A case study in Mali revealed that only around 35% of rainfall infiltrates the soil, whereas the rest is lost by surface runoff (Stroosnijder & Hoogmoed 1984 cited in Wallace & Batchelore 1997). Surface runoff is a problem for both rain-fed and irrigated agriculture, especially when rainfall is large and intensive, land is steep, or soil infiltration rates are low (Wallace 2000). A case study in India revealed that approximately 33% of rainfall was lost to percolation (Ong et al. 1991 cited in Wallace & Batchelore 1997). In West Africa, elevated water losses in rain-fed agriculture can be explained by a combination of infrequent but intense rainfalls such that crop transpiration is likely to be only between 15% and 30% of rainfall (Wallace 2000).

To increase water-use efficiency, several practices can be adopted. Runoff can be reduced by increasing surface storage, by leaving crop residues on the soil surface, or by planting cover crops and hedgerows (Lal 1989; Kiepe & Rao 1994). Direct evaporation of water from soil surface can be minimized by adopting adequate irrigation systems with higher field application efficiency, such as drip irrigation systems or adequate tillage practices. Hence, the improvement in the water-use efficiency should be the main focus in water-management policies instead of trying to supply the growing demand for water with new sources (Gleick 2003). However, getting more "crop per drop" involves a complex combination of investments as well as behavioral and institutional changes (World Bank 2010) that are frequently not in use.

Conclusions

An increase in the world's population and the food demand as well as the changing consumption patterns toward a higher proportion of livestock-based products are anticipated for future decades, posing challenges to agricultural production.

Globally, the amount of arable land remained relatively stable in the past. Nevertheless, agricultural land is being lost as a result of degradation processes and increasing competition from alternative uses (such as biofuel production, urbanization, or nature conservation and biodiversity protection). Intensification of agricultural production has played a major role in satisfying the global food demand in the past. The use of nitrogen and phosphorus fertilizers as well as the expansion of irrigated areas is predicted to increase on a global scale in the future. However, overuse of fertilizer, as well as irrigation, can have detrimental effects on the environment and ecosystems, which can impede food production in the future. Additionally, agricultural production might face water scarcity that could be aggravated by climate change in arid and semi-arid regions.

To relieve pressure on resources and environment as well as reduce the regional and global mismatch between food production and consumption, our suggestions aim at more efficient handling of resources and food including (1) changing consumption patterns and compositions between plant- and animal-based food components, (2) reducing food losses in the food supply chain such as postharvest losses or losses within households where eatable products are frequently discarded, and (3) sustainably intensifying agricultural production by using land, nutrients, and water more efficiently through better technologies. These measures might enable an increase in food supply without putting additional pressures on natural resources and the environment.

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